



# The Open Parasitology Journal

Content list available at: [www.benthamopen.com/TOPARAJ/](http://www.benthamopen.com/TOPARAJ/)

DOI: 10.2174/1874421401806010087, 2018, 6, 87-95



## RESEARCH ARTICLE

# Factors Affecting Helminth Abundances in Synanthropic Rodents of an Urban Environment

Diego Hancke<sup>1,2,\*</sup> and Olga Virginia Suárez<sup>1,2</sup>

<sup>1</sup>Laboratorio de Ecología de Roedores, Departamento de Ecología, Genética y Evolución, Facultad de Ciencias Exactas y Naturales, Universidad de Buenos Aires, Intendente Güiraldes 2160, Ciudad Universitaria, PB II, 4to piso, C1428EHA Ciudad Autónoma de Buenos Aires, Argentina

<sup>2</sup>Instituto de Ecología, Genética y Evolución de Buenos Aires (IEGEBA), UBA-CONICET, Intendente Güiraldes 2160, Ciudad Universitaria, PB II, 4to piso, C1428EHA Ciudad Autónoma de Buenos Aires, Argentina

Received: June 22, 2018

Revised: October 1, 2018

Accepted: October 8, 2018

### Abstract:

#### Background:

Current levels of urbanization cause changes in the ecology of hosts, the pathogens, or both, promoting the proliferation of zoonotic diseases. Rodents are a good biological model for the development of pathogen transmission models because its presence is often related to a none-adequate environmental management.

#### Objective:

The main goal of this paper was to study the changes in the abundance of helminth populations in synanthropic rodents of an urban landscape.

#### Methods:

A total of 92 *R. norvegicus* and 65 *M. musculus* were captured in the City of Buenos Aires (Argentina) and were screened for parasites. The variations in helminth abundances were studied at host population scale to determine the factors, such as the type of environment, meteorological conditions and demographic parameters of the hosts, which have an effect on helminth infection rates.

#### Results:

Parasites with intermediate hosts or free living larval stages in their life cycle were the most affected. It was found how rodents' use of the habitats in the different urban environments has an effect on the helminth infection levels. Besides, the importance of season on helminth abundance was determined, suggesting that climatic conditions are crucial for parasite survival and transmission.

#### Conclusion:

This information is relevant because it not only allows us to deepen the ecological dynamics of parasites in urban rodents, but also shows that environmental conditions are determinants for the persistence of helminth populations in a city.

**Keywords:** Helminths, Rodents, Urbanization, Ecology, Zoonosis, Buenos Aires.

## 1. INTRODUCTION

The current levels of human–ecosystem interaction result in a habitat transformation and changes in the ecology of

\* Address correspondence to the author at the Laboratorio de Ecología de Roedores, Departamento de Ecología, Genética y Evolución, Facultad de Ciencias Exactas y Naturales, Universidad de Buenos Aires, Avenida Intendente Cantilo s/n, Ciudad Universitaria, Pabellón II, 4º Piso Laboratorio 104 (C1428EHA), Buenos Aires, Argentina, Tel: +541145763300int219; Fax: +541145763354; E-mail: [diegohancke@ege.fcen.uba.ar](mailto:diegohancke@ege.fcen.uba.ar)

the host, the pathogen, or both, promoting thus an alteration in disease transmission dynamics [1, 2]. Urbanization is a phenomenon that accelerates and intensifies these impacts on zoonotic diseases [3, 4]. This is particularly observed for urban rodents because of the synanthropic behavior of these animals [5].

*Mus musculus* and *Rattus norvegicus* are among the best urban adapted rodents. Originated from Southeast Asia, various aspects of their biology, such as enormous reproductive potential, feeding benefited by accumulated or discarded human food, adaptations to urban environments, contribute to their success worldwide [6]. Synanthropic rodents are the main reservoirs or host for different human pathogens, including zoonotic helminth species [7]. The presence of murine rodents in anthropogenic habitats could be the reason for many of the parasitological studies performed worldwide [5], increasing the detection of rodent transmitted pathogens in last decades [8].

In the City of Buenos Aires (Argentina) (CBA), the community of rodents and their parasites has been studied for more than a decade. The establishment and proliferation of these animals are affected by the urban context, varying rodent composition according to habitat characteristics. *Rattus norvegicus* and *M. musculus* are the most common species and are the dominant species in two different landscape units of the CBA: shantytowns and parks or green spaces [9].

The helminth fauna of *R. norvegicus* and *M. musculus* of the CBA was recently described, exhibiting each rodent species its own characteristics in terms of richness, diversity and composition [10]. A high similarity was seen in the structure of the infracommunities of *R. norvegicus* from shantytowns and parklands, probably related to similar environmental characteristics that predict the presence of brown rats in both landscape units [11]. In contrast, *M. musculus* is preferably captured inside the houses in shantytowns, exploiting thus microhabitats with different conditions regarding the specimens captured in parks, which result in quantitative differences between infracommunities from both landscape units [11]. However, the factors affecting the abundance of helminth populations of urban rodents are still unknown.

Parasites have been considered as excellent bioindicators of processes occurring in an ecosystem because their life cycles depend on the presence of intermediate hosts or are involved in trophic chains or its survival and development occur in specific environmental physical-chemical ranges, among many others [8, 12]. The greatest abundances of each parasite species are thus expected in sites where the resources (including the hosts) are in full and the conditions are propitious for the transmission and survival of free life stages [13]. The main goal of this paper is to study the changes of the abundance in helminth populations of *R. norvegicus* and *M. musculus* captured in the CBA and analyze the factors that explain these variations. It is expected that helminths with indirect life cycles were more affected by external factors than those with direct cycles and covariations are expected to occur between parasite species with similar life cycles and resource requirements.

## 2. MATERIALS AND METHODOLOGY

The study was carried out in Buenos Aires (34°37 'S; 58°24 'W), the main city of Argentina in terms of population and one of the largest metropolises in the world [14]. Information about the study area, rodent collection and parasitological prospection is provided in detail in Hancke and Suárez (2017) [10]. For the purpose of this paper, a total of 157 rodents, 92 *R. norvegicus* and 65 *M. musculus*, captured in 4 shantytowns and 3 parklands were analyzed.

First, possible variations in the abundance between landscape units were analyzed for both rodent species. Abundance was estimated by trap success and calculated as the proportion of traps with captured animals out of the total number of trap-nights for each site in each trapping survey. The relationship between trap success of *R. norvegicus* and trap success of *M. musculus* was analyzed for shantytowns, parks and totally by linear regressions. For each rodent species, trap successes between shantytowns and parklands were compared by analyses of variances (ANOVA). Trap success was root square transformed to meet normality assumption.

In parasitological studies, it is recommended to define the limits of parasite populations or communities at different scales. According to Bush *et al.* (1997) [15]:

Infracommunities refer to a community of infrapopulations, which include all individuals of a parasite species in a single host at a particular time. Component community refers to all infrapopulations of parasites associated with a subset of hosts in a particular time and place (or in a given ecosystem). Prevalence, mean intensity (I) and abundance (A) were calculated to describe the distribution of each parasite species in both host species following as recommended by Bush *et al.* (1997) [15].

For each host species, Permutational Multivariate Analyses of Variance (PERMANOVA) [16] were performed to study the effects of trapping season and landscape unit on the composition of infracommunities. To test if helminth species are independent of each others, the adjustment of the distribution of infracommunity richness frequencies to the model of Janovy *et al.*, (1995) [17] was studied for both rodent species. This model assumes that, in the absence of associations and interactions among parasite species, the distribution of the observed frequencies of the infracommunity richness can be predicted by the prevalence values of each of the parasite species. Therefore, each species would have an associated probability of infection which would be independent of the presence of other species in that host. All distributions were tested using the Chi-Square goodness of fit test.

Zero-Inflated Negative Binomial models (ZINB) were performed to study the relationship between the abundance of the helminth infrapopulations and the characteristics of the hosts [18]. Only the species whose prevalence values were equal or greater than 10% were included. ZI distributions can be viewed as a two-part model, in which the probability of species presence and abundance are modeled from the same data [19]. Particularly in Zero-Inflated Negative Binomial (ZINB) models, the probability of a zero outcome is modeled by logistic regression, while the continuous count is modeled by using a negative binomial error structure. The separate occurrence and abundance terms can represent different mechanisms due to factors operating at different temporal or spatial scales [19]. The number of worms per rat was considered as the dependent variable, whereas factors such as landscape units, season, sex, year, body length (covariable), abundance of *R. norvegicus* (covariable) and abundance of *M. musculus* (covariable) were tested in both the logistic and the negative binomial parts of the model as potential independent variables. The ZINB models were fitted manually, testing the significance of factors, covariable and interactions, first of the binomial part and then of the count part. The models were then progressively simplified by backward deletion and only significant terms ( $P < 0.05$ ) were left in the final models. The models were selected with likelihood ratio tests and Akaike information criteria (AIC). For the particular cases of helminth species with low-intensity values, in this case, it makes more sense to evaluate the presence and not abundance, generalized linear models assuming a distribution of binary binomial errors or Bernoulli (presence / absence) with logit link function were conducted.

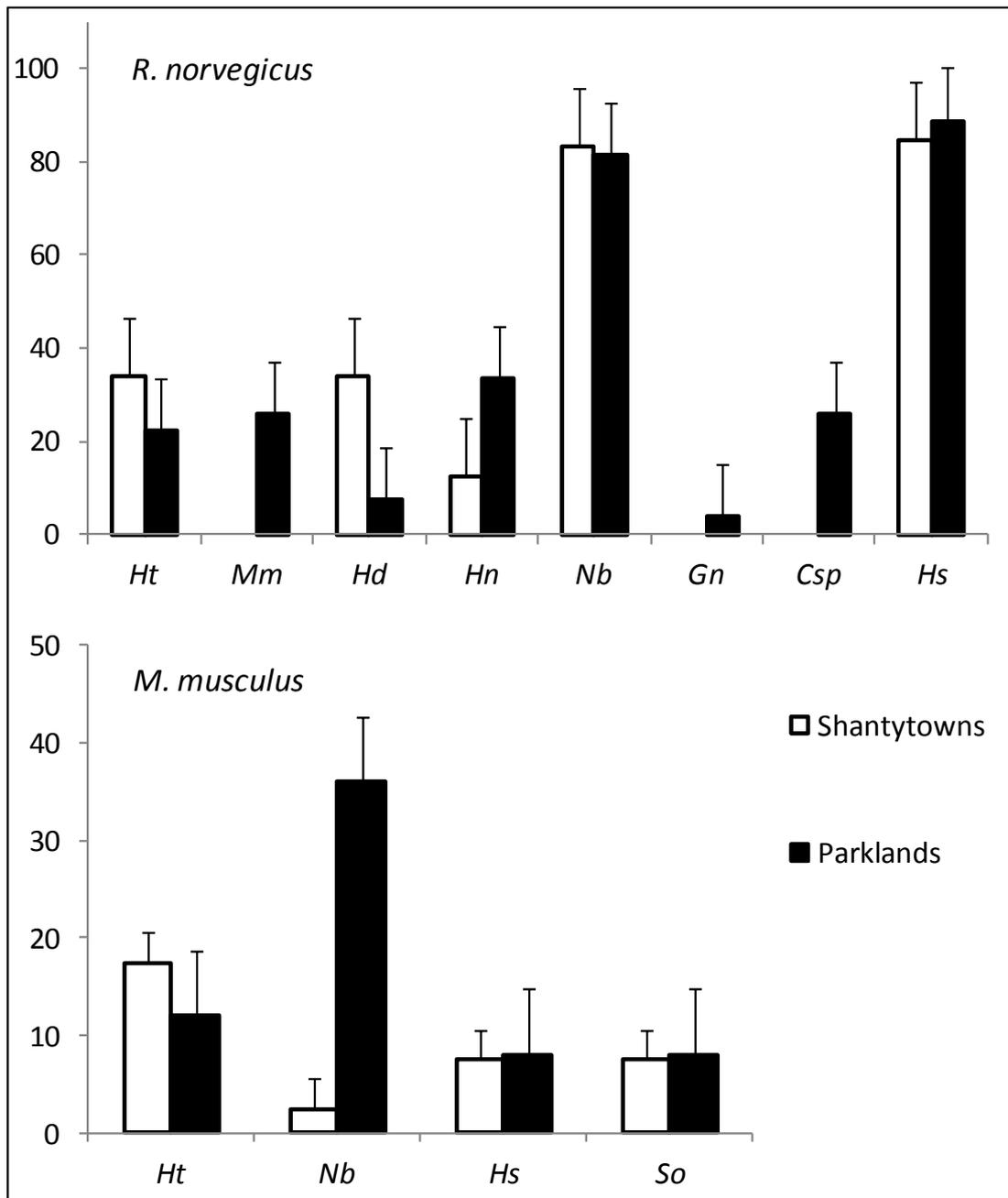
The relationship between the abundances of the most representative helminth species in each host species was tested in pairs with ZINB models. In each case, the abundance of one of the helminth species was considered as a dependent variable and the other as explanatory variable. As in the previous ZINB models, each host was considered as the experimental unit. Given the small abundances of *Hymenolepis nana* and *Hydatigera taeniaeformis*, the presence instead of abundance was used as an explanatory variable. As the abundance of parasite can vary between environments, the landscape unit was included as a fixed factor.

All calculations were performed using the R version 2.15.1 [20]. PERMANOVA analysis was performed using the vegan package [21]. ZINB models were fitted with pscl package [22] and all likelihood ratio tests were carried out in lmttest package [23].

### 3. RESULTS

The abundances (or trap success) for *R. norvegicus* and *M. musculus* were independent of each other considering the total of the sites (slope=0.35, 95% CI [-0.72, 1.42],  $p > 0.05$ ,  $R^2 = 0.02$ ) as well as considering shantytowns only (slope=-0.94, 95% CI [-3.09, 1.21],  $p > 0.05$ ,  $R^2 = 0.08$ ). For parklands, a positive relationship was observed between the trapping success of both rodent species (slope=1.43, 95% CI [0.44, 2.48],  $p < 0.05$ ,  $R^2 = 0.50$ ). While comparing abundance of the rodent, a higher trap success was detected in *M. musculus* in shantytowns (ANOVA,  $N = 25$ ,  $F = 8.37$ ,  $p < 0.05$ ) while no differences were observed for *R. norvegicus* (ANOVA,  $N = 25$ ,  $F = 0.39$ ,  $p > 0.05$ ).

*Rattus norvegicus* was parasitized with a total of 8 helminth species (8 in parklands and 5 in shantytowns) while *M. musculus* harboured the same 4 species in both landscape units. Fig. (1) shows variations in the prevalence of these species between landscape units for both hosts. A more detailed information about the infection parameters is provided in Hancke and Suárez (2017) [11]. The PERMANOVA analysis revealed that the variations in the composition of infracommunities varied differently in *R. norvegicus* and in *M. musculus* Table 1. An interaction was observed between landscape units and the time of capture in brown rats. The effect of the time of capture was thus analyzed for each landscape unit separately and a significant effect was detected only for rats captured in parklands (shantytowns: PERMANOVA, Pseudo-F = 2.098,  $p > 0.05$ ; parks: PERMANOVA, Pseudo-F = 2.971,  $p < 0.05$ ). For *M. musculus*, variations in the composition of infracommunities were detected between landscape units and between capture seasons Table 1.



**Fig. (1).** Prevalence (%) (with std errors) for the helminth species of *R. norvegicus* and *M. musculus* from shantytowns and parklands of the City of Buenos Aires. Acantocephalans: Mm – *Moniliformis moniliformis* (Bremsler, 1811); Cestodes; Ht – *Hydatigera taeniaeformis* (Batsch, 1786); Hd – *Hymenolepis diminuta* (Rudolphi, 1819); Hn – *Hymenolepis nana* (Siebold, 1852); Nematodes; Gn – *Gongylonema neoplasticum* (Fibiger et Ditlevsen, 1914); Nb – *Nippostrongylus brasiliensis* (Travassos, 1914); Csp – *Capillaria* sp.; Hs – *Heterakis spumosa* Schneider, 1866.

**Table 1.** Output of PERMANOVA analysis examining the helminth composition variations between landscape units and trapping seasons in *R. norvegicus* and *M. musculus* captured in the City of Buenos Aires.

Factor	df	SS	MS.	Pseudo-F.	P
<i>R. norvegicus</i>					
Landscape	1	0.30	0.30	1.21	0.297
Time	1	0.40	0.40	1.62	0.163
Lansc x Time	1	0.77	0.77	3.12	0.016

(Table 1) contd.....

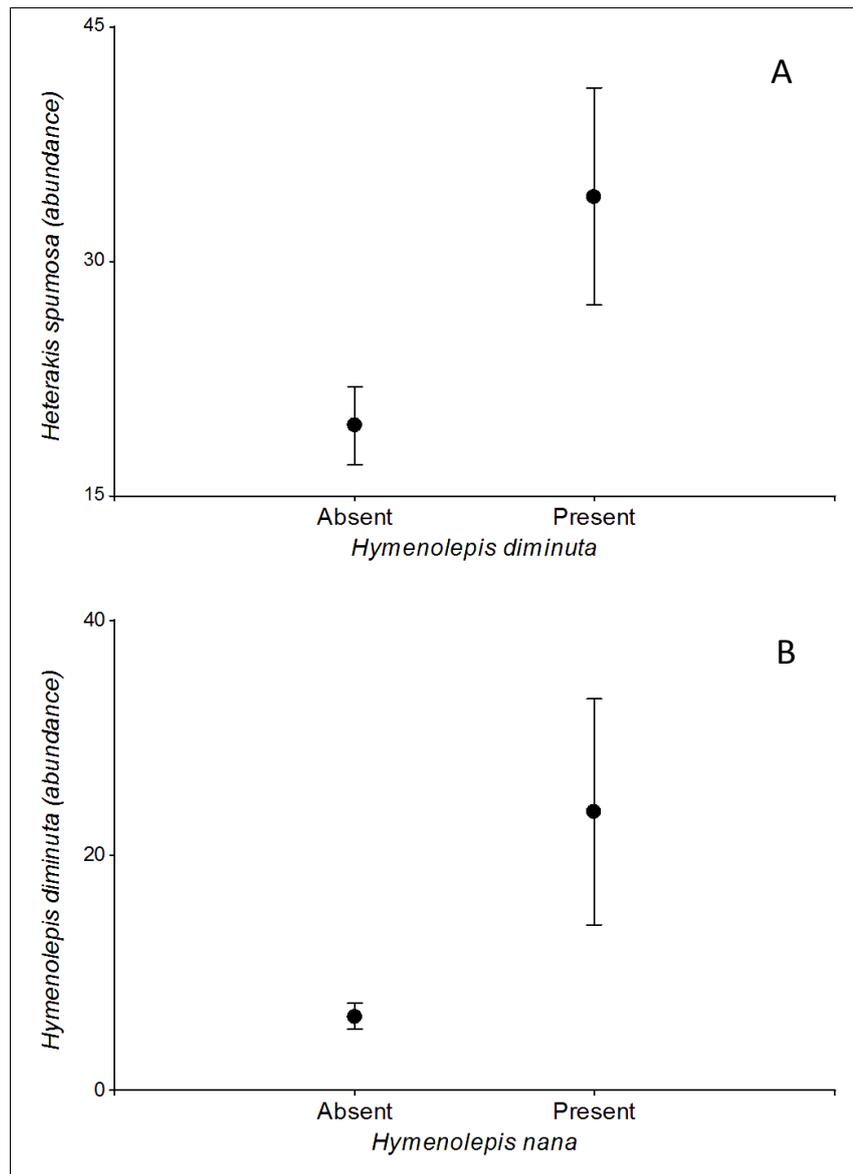
Factor	df	SS	MS.	Pseudo-F.	P
<i>R. norvegicus</i>					
Residuals	88	21.70	0.25		
Total	91	23.17			
<i>M. musculus</i>					
Landscape	1	0.29	0.29	2.22	0.057
Time	1	0.54	0.54	4.20	0.001
Lansc x Time	1	0.04	0.04	0.34	0.876
Residuals	61	7.86	0.13		
Total	64	8.73			

The distributions of the frequencies of species richness for both hosts did not differ from those expected by the Janovy model (*R. norvegicus*:  $\chi^2 = 0,009$ ; gl= 5; p= 0,999 (shantytowns),  $\chi^2 = 0,334$ ; gl= 8; p = 0,999 (parks); *M. musculus*:  $\chi^2 = 0,007$ ; gl= 4; p= 0,999 (shantytowns),  $\chi^2 = 0,025$ ; gl= 4; p = 0,999 (parks)). This means that the probability of infection by a single helminth species remains independent of the other species of an infracommunity. Table 2 summarizes the results of the generalized linear models for the factors affecting the helminth abundance. In the case of *R. norvegicus*, the type of landscape unit affected 3 species. On one hand, shantytowns had a positive effect on the presence of *Hymenolepis diminuta* and on the abundance of *Nippostrongylus brasiliensis*, while on the other side, parks had it on the presence of *H. nana*. Besides, the colder trapping season affected negatively the presence of *H. diminuta* and *H. nana* and negatively the abundance of *H. spumosa*. Both cestodes, *H. diminuta* and *H. nana*, were also affected by host gender and body length of *R. norvegicus*, while the abundance of *N. brasiliensis* and *H. spumosa* in rats were negatively and positively associated with the trap success of *M. musculus* respectively. In the case of *M. musculus*, the abundance of *N. brasiliensis* was positively affected by the colder trapping season and the trapping success of *R. norvegicus* and negatively with the abundance of *M. musculus*. *Hydatigera taeniaeformis* was analyzed for both rodent species, but no significant associations were found.

**Table 2. Summary of the results of the generalized linear models performed to analyze the factors that have an effect on the presence and abundance of helminths in *R. norvegicus* and *M. musculus* of the City of Buenos Aires (significant codes: “.” 0.05<p<0.1; “\*” 0.01<p<0.05; “\*\*\*” 0<p<0.01) (effect: “+” positive; “-” negative).**

Parasite	Variable	Predictor	Coeff.	Std. Err.	P	Effect	Life cycle
<i>R. norvegicus</i>							
<i>H. diminuta</i>	presence	Shantytown	-1.674	0.864	.	+	Indirect. Rodents as definitive. Arthropods as intermediate hosts.
		Cold	3.653	1.409	**	-	
	abundance	Male	0.919	0.474	.	+	
<i>H. nana</i>	presence	Shantytown	-1.216	0.588	*	-	Direct or Indirect. Arthropods as intermediate hosts.
		Cold	-1.018	0.593	.	-	
		Lengh	0.022	0.013	.	+	
<i>N. brasiliensis</i>	abundance	Shantytown	0.691	0.320	*	+	Geohelminth
		Ab. Mm	-0.086	0.0376	*	-	
<i>H. spumosa</i>	abundance	Cold	-0.621	0.315	*	-	
		Ab. Mm	0.072	0.040	.	+	
<i>M. musculus</i>							
<i>N. brasiliensis</i>	Presence	Cold	3.817	1.462	**	+	-
		Ab. Mm	-0.608	0.187	**	-	-
		Ab. Rn	0.363	0.169	*	+	-

Regarding the interspecific helminth covariations, only significant effects were found on the abundance of *H. diminuta*. On one hand, the probability of infection of this cestode was positively associated with the abundance of *H. spumosa* (p <0.05), while on the other hand, its abundance increased with the presence of *H. nana* (p <0.05) (Fig. 2).



**Fig. (2).** Graphs representing the mean abundance of *H. spumosa* of infracommunities of *R. norvegicus* with and without *H. diminuta*(A) and the mean abundance of *H. diminuta* of infracommunities with and without *H. nana*(B). Bars represents the standard errors.

#### 4. DISCUSSION

According to previous studies, helminth infracommunities of urban rodents could be grouped according to composition and relative abundances and they respond to the structure of the host community [11]. In this paper, we analyzed the variations in helminth abundances at a host population scale and identified the factors affecting the principal parasite species. Rodents are the main reservoirs or hosts of parasites relevant for public health. In this study, the presence of *H. nana* and *H. diminuta* confirms this role. Rodents represent a good biological model for the development of pathogen transmission models in urban landscapes because its presence is often a consequence of none-adequate environmental managements.

The type of urban landscape unit had a significant effect on the two intestinal cestodes found in our study. On one hand, a higher prevalence of *H. diminuta* was observed in shantytowns, showing that the conditions here favor the proliferation of arthropod pests as *Tenebrio castaneum* and *T. confusum*, both intermediate hosts of *H. diminuta* [24]. On the other hand, parks favor the presence of *H. nana*. This parasite is the most common cestode found in humans around the world, especially in children from poor areas [25, 26]. The parks sampled in our study area were located on

the banks of rivers on whose shorelines precarious settlements not far from the parks take place. This situation would favor a greater organic matter contamination which could act as a source of intestinal parasite cysts or eggs, among other pathogens. However, despite both species were favored by different urban habitats at a landscape scale, a positive association between both species was observed. *Hymenolepis nana* is the only known cestode species that presents a direct cycle and occasionally uses arthropods as intermediate hosts, the same as *H. diminuta*. It could be hypothesized that both species may be responding at a microscale to similar biotic and abiotic factors, mediating also intermediate hosts the transmission of *H. nana* in *R. norvegicus*.

This is in concordance with the fact that both cestodes were also affected by the time of capture, decreasing both prevalences in the colder months. The development of intermediate host's life cycle depends on optimal temperature and humidity conditions, affecting thus the abundance of the parasite [27]. Warmer months favor also the abundance of *H. spumosa* in *R. norvegicus*. Season is mentioned to affect the behavior of rats within urban environments, which could produce seasonal variations in the infection parameters of parasite populations [28]. The fact that *H. diminuta* and *H. spumosa* were both favored by warmer conditions could explain the positive relationship observed between both species. Besides, the time of capture also affected *N. brasiliensis* in *M. musculus*, but in this case, colder months favored the presence of this nematode. Seasonal variations in *N. brasiliensis* have been previously reported but for *R. norvegicus* captured in a shantytown of CBA [29]. As *N. brasiliensis* is a soil transmitted helminth and whose larval stages require adequate soil moisture conditions and are not resistant to desiccation, its surveillance in the environment is possibly conditioned in the warmer months of the year [30].

Host density and body length were also mentioned to be important factors that influence helminth infection rates, especially for directed transmitted parasite [31]. In our study, the infection likelihood of *N. brasiliensis* in mice was favoured by greater densities of *R. norvegicus*. The abundance of this nematode was high in *R. norvegicus* (over 80%), so the infection of *N. brasiliensis* in *M. musculus* is possibly related to the presence of brown rats, particularly in parks where the abundance of both rodents co-varied. Besides, the negative effect of mice abundance on the presence of *N. brasiliensis* is probably masking an effect of landscape unit more than an effect of population density itself, because a significant lower trap success for *M. musculus* was detected in parks.

Spatial variations in infection rates have already been described for a wide range of intestinal helminth communities for rodents in urban environments [12, 24, 32]. In our study area, while analyzing the variations of infracommunities in terms of composition, *M. musculus* exhibited differences between shantytowns and parklands. *Mus musculus* in both landscape units make use of different types of microhabitats because in shantytowns, it was trapped mainly inside the housings (in contrast to *R. norvegicus* who was captured mainly outside). For *R. norvegicus*, no differences were detected between the composition of infracommunities from shantytowns and parks, because the landscape unit did not affect the presence of the core species [10]. But an effect of the season was observed but only for rats captured in the parks. This suggests that seasonal fluctuations of the conditions for parasite survival and transmission may be reduced in the more human-dominated habitats like shantytowns.

## CONCLUSION

To conclude, although host identity is a strong predictor of the structure of helminth infracommunities (diversity, composition, core species) within an urban rodent assembly [11], helminth infection rates are affected by external factors such as the type of environment, meteorological conditions and demographic parameters of the hosts. Particularly, parasites with intermediate hosts or free living larval stages in their life cycle are the most affected species. This information is relevant because it allows us to deepen the ecological dynamics of parasites in urban rodents and shows that environmental conditions are determinants for the persistence of urban helminth populations. The link between pathogens-rodent-environment is intimately connected, needing holistic approaches to address the prevention of zoonoses. Therefore, an improvement of the hygienic and environmental quality must be considered in any health program against urban zoonoses, particularly in shantytowns where social, economical and environmental conditions expose their inhabitants to a high risk of the so-called neglected diseases.

## ETHICS APPROVAL AND CONSENT TO PARTICIPATE

Not applicable.

## HUMAN AND ANIMAL RIGHTS

No animals/humans were used for studies that are the basis of this research.

## CONSENT FOR PUBLICATION

Not applicable.

## CONFLICT OF INTEREST

The authors declare no conflict of interest, financial or otherwise.

## ACKNOWLEDGEMENTS

The authors wish to thank all the team of the Laboratorio de Ecología de Roedores Urbanos for their assistance during the field sampling. We are also grateful to Dra. Graciela Navone (CEPAVE-CONICET-La Plata-Argentina) for her help provided during helminth identification and her valuable comments on earlier versions of the manuscript. Financial support was provided by the University of Buenos Aires, Gobierno de la Ciudad de Buenos Aires and CONICET (Argentina).

## REFERENCES

- [1] Thompson RC. Parasite zoonoses and wildlife: One Health, spillover and human activity. *Int J Parasitol* 2013; 43(12-13): 1079-88. [<http://dx.doi.org/10.1016/j.ijpara.2013.06.007>] [PMID: 23892130]
- [2] Hassell JM, Begon M, Ward MJ, Fèvre EM. Urbanization and disease emergence: Dynamics at the wildlife-livestock-human interface. *Trends Ecol Evol (Amst)* 2017; 32(1): 55-67. [<http://dx.doi.org/10.1016/j.tree.2016.09.012>] [PMID: 28029378]
- [3] Mackenstedt U, Jenkins D, Romig T. The role of wildlife in the transmission of parasitic zoonoses in peri-urban and urban areas. *Int J Parasitol Parasites Wildl* 2015; 4(1): 71-9. [<http://dx.doi.org/10.1016/j.ijppaw.2015.01.006>] [PMID: 25830108]
- [4] Neiderud CJ. How urbanization affects the epidemiology of emerging infectious diseases. *Infect Ecol Epidemiol* 2015; 5: 27060. [<http://dx.doi.org/10.3402/iee.v5.27060>] [PMID: 26112265]
- [5] Morand S, Bordes F, Chen HW, *et al.* Global parasite and Rattus rodent invasions: The consequences for rodent-borne diseases. *Integr Zool* 2015; 10(5): 409-23. [<http://dx.doi.org/10.1111/1749-4877.12143>] [PMID: 26037785]
- [6] Kosoy M, Khlyap L, Cosson JF, Morand S. Aboriginal and invasive rats of genus *Rattus* as hosts of infectious agents. *Vector Borne Zoonotic Dis* 2015; 15(1): 3-12. [<http://dx.doi.org/10.1089/vbz.2014.1629>] [PMID: 25629775]
- [7] Himswoth CG, Parsons KL, Jardine C, Patrick DM. Rats, cities, people, and pathogens: A systematic review and narrative synthesis of literature regarding the ecology of rat-associated zoonoses in urban centers. *Vector Borne Zoonotic Dis* 2013; 13(6): 349-59. [<http://dx.doi.org/10.1089/vbz.2012.1195>] [PMID: 23590323]
- [8] Bordes F, Blasdell K, Morand S. Transmission ecology of rodent-borne diseases: New frontiers. *Integr Zool* 2015; 10(5): 424-35. [<http://dx.doi.org/10.1111/1749-4877.12149>] [PMID: 26176684]
- [9] Cavia R, Cueto GR, Suárez OV. Changes in rodent communities according to the landscape structure in an urban ecosystem. *Landsc Urban Plan* 2009; 90: 11-9. [<http://dx.doi.org/10.1016/j.landurbplan.2008.10.017>]
- [10] Hancke D, Suárez OV. Helminth diversity in synanthropic rodents from an urban ecosystem. *EcoHealth* 2017; 14(3): 603-13. [<http://dx.doi.org/10.1007/s10393-017-1239-8>] [PMID: 28417211]
- [11] Hancke D, Suárez OV. Structure of parasite communities in urban environments: The case of helminths in synanthropic rodents. *Folia Parasitologica* 2018; 65(1): 2018-009. [<http://dx.doi.org/10.14411/fp.2018.009>] [PMID: 30183669]
- [12] Froeschke G, Matthee S. Landscape characteristics influence helminth infestations in a peri-domestic rodent--implications for possible zoonotic disease. *Parasit Vectors* 2014; 7: 393. [<http://dx.doi.org/10.1186/1756-3305-7-393>] [PMID: 25159989]
- [13] Lagrue C, Poulin R. Spatial covariation of local abundance among different parasite species: The effect of shared hosts. *Parasitol Res* 2015; 114(10): 3637-43. [<http://dx.doi.org/10.1007/s00436-015-4590-0>] [PMID: 26113509]
- [14] United Nations. World Urbanization Prospects: The 2014 Revision. Highlights 2014.
- [15] Bush AO, Lafferty KD, Lotz JM, Shostak AW. Parasitology meets ecology on its own terms: Margolis *et al.* revisited. *J Parasitol* 1997; 83(4): 575-83. [<http://dx.doi.org/10.2307/3284227>] [PMID: 9267395]
- [16] Anderson MJ. A new method for non-parametric multivariate analysis of variance. *Austral Ecol* 2001; 26: 32-46.

- [17] Janovy J Jr, Clopton RE, Clopton DA, Snyder SD, Efting A, Krebs L. Species density distributions as null models for ecologically significant interactions of parasite species in an assemblage. *Ecol Modell* 1995; 77: 189-96. [[http://dx.doi.org/10.1016/0304-3800\(93\)E0087-J](http://dx.doi.org/10.1016/0304-3800(93)E0087-J)]
- [18] Zuur AF, Ieno EN, Walker NJ, Saveliev AA, Smith GM. *Mixed effects models and extensions in ecology with R*. New York: Springer 2009. [<http://dx.doi.org/10.1007/978-0-387-87458-6>]
- [19] Wenger SJ, Freeman MC. Estimating species occurrence, abundance, and detection probability using zero-inflated distributions. *Ecology* 2008; 89(10): 2953-9. [<http://dx.doi.org/10.1890/07-1127.1>] [PMID: 18959332]
- [20] R Development Core Team. *R: A language and environment for statistical computing*. R Foundation for Statistical Computing 2014. Available from: <http://www.R-project.org/>
- [21] Oksanen AJ, Blanchet FG, Kindt R. Package ‘vegan’ version 2.5-2. 2018
- [22] Maimone C, Fearon J, Jackman MS. Package ‘pscl.’ 2012.
- [23] Hothorn T, Zeileis A, Farebrother RW, Cummins C, Mollo G, Mitchell D. Package “lmtest”. *Testing Linear Regression Models* 2018.
- [24] Hancke D, Suárez OV. Infection levels of the cestode *Hymenolepis diminuta* in rat populations from buenos aires, argentina. *J Helminthol* 2016; 90(2): 199-205. [<http://dx.doi.org/10.1017/S0022149X15000164>] [PMID: 25869333]
- [25] Roberts LS, Janovy J. *Foundations of Parasitology*. 8th ed. New York: McGraw-Hill 2009.
- [26] Malheiros AF, Mathews PD, Lemos LMS, Braga GB, Shaw JJ. Prevalence of *Hymenolepis nana* in indigenous Tapirapé ethnic group from the brazilian amazon. *Am J Biomed Res* 2014; 2: 16-8. [<http://dx.doi.org/10.12691/ajbr-2-2-1>]
- [27] Robinson WH. *Urban insects and arachnids: A handbook of urban entomology*. Cambridge University Press 2015.
- [28] Himsforth CG, Jardine CM, Parsons KL, Feng AYT, Patrick DM. The characteristics of wild rat (*Rattus* spp.) populations from an inner-city neighborhood with a focus on factors critical to the understanding of rat-associated zoonoses. *PLoS One* 2014; 9(3): e91654. [<http://dx.doi.org/10.1371/journal.pone.0091654>] [PMID: 24646877]
- [29] Hancke D, Navone GT, Suárez OV. Endoparasite community of *Rattus norvegicus* captured in an urban area of Argentina. *Helminthologia* 2011; 48: 167-73. [<http://dx.doi.org/10.2478/s11687-011-0025-3>]
- [30] Anderson RC. *Nematode parasites of vertebrates: Their development and transmission*. New York: Cabi Publishing 2000. [<http://dx.doi.org/10.1079/9780851994215.0000>]
- [31] Arneberg P. Host population density and body mass as determinants of species richness in parasite communities: Comparative analyses of directly transmitted nematodes of mammals. *Ecography* 2002; 25: 88-94. [<http://dx.doi.org/10.1034/j.1600-0587.2002.250110.x>]
- [32] Zain SNM, Behnke JM, Lewis JW. Helminth communities from two urban rat populations in Kuala Lumpur, Malaysia. *Parasit Vectors* 2012; 5: 1-23. [PMID: 22212459]

---

© 2018 Diego and Virginia.

This is an open access article distributed under the terms of the Creative Commons Attribution 4.0 International Public License (CC-BY 4.0), a copy of which is available at: <https://creativecommons.org/licenses/by/4.0/legalcode>. This license permits unrestricted use, distribution, and reproduction in any medium, provided the original author and source are credited.